



***Nigella sativa* L. Seeds Biomass as a Potential Sorbent in Sorption of Lead from Aqueous Solutions and Wastewaters**

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ABSTRACT

This first study aims to establish the adsorption capacity of *Nigella sativa* seeds, an agro-food waste towards Pb(II) according to batch mode by varying contact time, initial Pb(II) concentration, initial pH and sorbent dose. The amount of Pb(II) ion adsorbed at interface is found high in order of 90.31 mg/g under optimum conditions. The FTIR characterization indicates the involvement of the hydroxyl and carbonyl groups of the given material in the adsorption of Pb(II). Isotherm curves and kinetic studies demonstrate that the Langmuir model and pseudo-second order kinetic describe well the sorption phenomenon and fit best to the sorption experimental data. Removing trial of lead from activated sludge waste and its decanted effluent confirm the efficiency of this biomass, which attain 97% and 64% respectively. Based on the results, the *Nigella sativa* seeds waste biomass appears as a potential eco-friendly and low cost sorbent for sorption of lead from wastewaters.


Keywords: *Nigella* seeds meals waste, Biosorption, Biomass sorbent, Heavy metal pollutants, Lead removal, Sorption isotherm

INTRODUCTION

In the field of environmental protection, removal of heavy metal pollutants from solid or liquid effluents constitutes a great challenge for health and environment institutions particularly in developing countries as a result of increasing industrial activities. Lead is one of the most used

metals in batteries accumulators manufacturing and metal extraction and finishing activities. It can be considered as one of the most toxic heavy metals that have latent long-term negative impacts on health, causing anemia, encephalopathy, hepatitis and nephritic syndrome¹. Therefore, there is urgent need for decontaminating these effluents before discharge them in the environment. Different



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technologies of treatment have been used and evaluated in removing toxic metals from industrial effluents based on various chemical or physical principles including precipitation by hydroxides, carbonates or sulfides², adsorption on kaolinite and montmorillonite³, membrane filtration⁴ and electrochemical methods such as electrocoagulation and electrowinning⁵. In last decade, biosorption process which uses various materials of biological origin, including bacteria, fungi, yeast and algae as bioadsorbents constitutes an alternative way in removing of metal pollutants^{6, 7, 8}. Waste products from food industry and agricultural activities are another emerged technology and innovative eco-friendly solution to the problem of heavy metal pollution due their abundance, cost-effective and high efficiency in removing ions pollutants and without secondary effect on environment⁹. In fact, some achieved investigations on various cheap bioadsorbents, non living plant material, such as potato peels¹⁰, eggshells¹¹, orange, lemon and banana peels^{12, 13}, coffee husks¹⁴, rice hull¹⁵, cotton and maize stalks and rice straw¹⁶ possess better adsorption capacities for a large spectrum of heavy metals depending on their chemical composition and operating conditions. For example, it was demonstrated that the efficiency removal of cadmium, iron and zinc by rice hull is higher than sawdust, sugarcane bagasse and wheat straw in the order of 98,15%, 96,90%, 93,00% and 91,19% respectively¹⁶. The high heavy metals sequestration of rice hull than others agriculture waste biosorbents was probably due to the presence of silanol groups in its structure as the authors suggested. In the study achieved by Mosa *et al.*,¹⁷ it was established that the adsorptive power of heavy metals decreases in the order of cotton stalks, maize stalks and rice straw. They attributed the potent removing efficacy of cotton stalks to its highest content of cellulose, hemicellulose and lignin. *Nigella sativa* L. (*Nigella*) is an annual herbaceous plant belonging to the ranunculaceae family growing in Asia, in northern Africa and in southern Europe. The black seeds of *Nigella*, known in Arabia as 'Al-Habba Al-Sawdaa' black cumin or Habatul-Barakah, have long been used as spice in food and in folk medicine in the worldwide particularly in the Middle and Far East for a wide range of diseases^{18, 19, 20}. The chemical composition of the seeds of *Nigella* has been largely investigated^{21, 22, 23}. The analysis of whole mature

seeds from different geographic and climatic origins showed that the crude protein vary from 20% to 27%, carbohydrates from 23.5% to 40, 0% and total lipid from 28, 48% to 40.35%. Polyphenols a secondary metabolite constitute another compound of this seeds with a high content in order of 27, 8 mg equivalent acid gallic per gram of extract²⁴. In addition, the oily fraction of the seeds is the active ingredients especially the essential oils^{25, 26} that possess a potential multi-purposes product largely used in phytotherapy and alternative medicine. For this reasons, oil factory was grown in Algeria, which engender, every year, several hundreds of tons of *Nigella* seeds meals waste that were discarded in nature as a fertilizer or used only for animal breeding as a complement. The aim of this study, the first at our knowledge, consists to establish the adsorptive properties of the biomass of *Nigella* seeds meals waste as bioadsorbent toward Pb (II), a divalent metal ion that is predominantly present in the effluents of battery factory and lead refining unit implanted at Sétif (East part region of Algeria). The adsorption approach was conducted in batch by using synthetic aqueous solutions at different contact time by varying the concentration of this metal at various pH using different biomass sorbent dose. The adsorption isotherms and sorption kinetics models of Pb(II) ions onto *N. sativa* seeds waste were also established. Fourier transform infrared spectroscopy (FTIR) was used to elucidate the adsorption process. Finally, the adsorption capacities was carried out by using a given amount of this bulk biomass in removing lead from activated sludge and liquid effluent from factory.

MATERIALS AND METHODS

Materials

All the necessary chemicals are of analytical grade and obtained from Sigma–Aldrich (USA). Stock solution (1000 mg/L), which was prepared by dissolving Pb(NO₃)₂ in deionized water. The sample solutions were prepared at the required concentrations by serial dilution of stock solution for practical use. The pH value of the solution was measured with a digital pH meter (AR 25 Dual Channel Fisher Scientific, Canada) using solid electrode calibrated with a standard buffer solution. The pH value of the solutions was adjusted by addition of 0.1 M HCl or NaOH. Atomic absorption

spectrophotometer (Perkin Elmer PinAA-900F, Canada) was used to assess the Pb(II), before and after biosorption, content in aqueous solution. FT-IR spectrophotometer (Shimadzu, Algeria) in the range of 500–4000 cm^{-1} was used to identify the functional groups of seed biomass powder, preliminarily dispersed in KBr pellets, before and after sorption process.

METHODS

Preparation of biosorbent

Firstly, the seeds of *Nigella* obtained from a local market of Tamalous (East part of Skikda, Algeria) during the end of 2015 were pre-treated to remove the surface impurities and then crushed into fine powder by mechanical grinder. The powder was sieved to an average size of 200–500 μm in order to allow a good mass transfer between sorbat and biosorbent. The *Nigella* seeds meals waste biomass was prepared by defatted action, which was achieved by organic solvent instead mechanic system as it done in oil factory. For that, an amount of powder was soaked in hexane in a ratio of 1/5 weight per volume for 2 h under agitation. This operation was repeated until total lipids were completely removed. Then, the biosorbent was filtered and dried in an incubator at 40 °C for 24 h for complete solvent removal. Finally, material was stored in an air tight bottle to protect it from contaminants and humidity until it uses as a biosorbent for the removal of Pb(II) experiments.

Batch adsorption experiments were carried out in flasks by adding 0.07 g of *N. sativa* seeds biomass in 10 ml of aqueous Pb(II) solution at desired initial pH (2, 4, 6, 8 and 10), initial metal ion concentration (93, 764, 1190, 1400 and 1819 mg/L) and contact time (90, 720, 1140 and 2880 minutes). The contents were shaken at 200rpm on at room temperature. After completion of each batch, the solution was separated by filtration with Whatman filter paper 41 followed by centrifugation at 3000rpm for 10 min. and the filtrate was analyzed for the remaining metal concentration in the sample using atomic absorption spectrophotometer. The amount of metal ion sorbed per unit mass of the biosorbent (mg/g) was evaluated by using the following equation 1,

$$Q = \frac{(C_0 - C_e)V}{m} \quad 1$$

The biosorption percentage (BS) of Pb(II) ions was calculated using the equation 2,

$$\text{BS \%} = \frac{(C_0 - C_e)}{C_0} \times 100 \quad 2$$

Where Q (mg/g) is the adsorption capacity, C_0 and C_e denote respectively the initial and equilibrium concentrations of metal ion (mg/L), V (L) is volume of adsorbate and m (g) is the amount of adsorbent. All the experiments were performed in duplicate and the results are the average values.

Characterization of biosorbent by FTIR

FTIR spectroscopy is a well known times saving method to characterize and identify functional groups²⁷ (Grube *et al.*, 2008). Therefore, FTIR spectrum was used as a tool to identify the characteristic functional groups that might be responsible for binding of metal ions. The FTIR spectra of before treatment and Pb(II) loaded *N. sativa* seeds were recorded in the range of 4000-500 cm^{-1} by pellet KBr method.

Trial removal of lead on real effluents

The *N. sativa* seeds biosorbent was tested for the removal of Pb(II) from wastewater effluents using predetermined optimum biosorption conditions. Material was brought from an activated sludge tank located in wastewater treatment plant in St. Hyacinthe, QC, Canada. The initial concentration of the mix liquor suspended solids was around 4000 mg/L. The tests were conducted on two different samples, i) directly on the activated sludge, and ii) on the decanted effluent of the activated sludge preliminarily centrifuged at 3000rpm and filtrated through a filter paper Whatman 41. Since the activated sludge contained traces of some metals only, the samples were adjusted to the level of Pb(II) to 764 mg/L. The pH value was not adjusted, and tests were conducted in neutral conditions (pH 6.35). An amount of 0.07 g of *N. sativa* seeds biomass were added to 10 ml of both samples and were exposed to 360 min. of contact time at ambient temperature. After treatment, the lead content in supernatants of both samples was determined by atomic absorption spectrophotometry

RESULTS AND DISCUSSION

Effect of the initial metal ions concentration

The sorption capacity (Q mg/g) of lead onto *N. sativa* seeds biomass increases from 7.79 to 39, 7 mg/g, when the initial Pb(II) ions concentration in medium increases from 93 to 1819 mg/L respectively (Fig. 1 A). This common finding and phenomenon could be probably due to a full contact or collision between sorbat and sorbent as established by Wang and Chen⁹. Inversely, the removal efficacy decreases by increasing the initial metal ion concentration. In fact, an increase of Pb(II) concentration in the same range result in a decrease of the removal efficiency from 99, 4 to 42,13%. At low concentration, most of the metal ions in solution interact with active sites on the biosorbent resulting in high percentage removal until the saturation plateau was attained^{28, 29}. However, at higher concentrations the diffusion of metal ions into biomass surface was achieved via interparticle interactions. The difference between bulk and surface metal ions concentration is one of the driving forces to overcome the resistance to adsorption in the absence of the mass transfer resistance surface and bulk concentrations²⁸. This is due to high initial Pb(II) ion concentration, where the number of moles of metal ions available to the surface area might be high, therefore functional adsorption becomes dependent on the initial metal ion concentration²⁸. This result is comparable with to those reported in bibliography for other adsorbents^{30, 31}. Consequently, the initial concentration of lead of 764 mg/L was used in subsequent tests in this study.

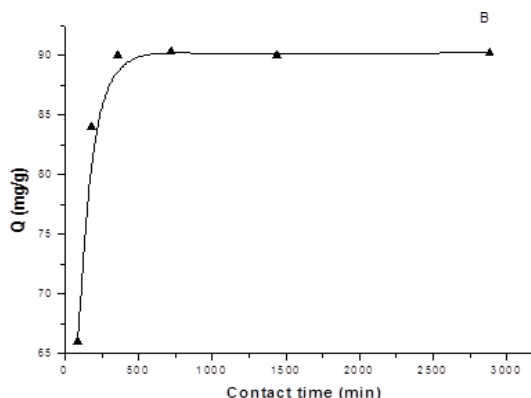
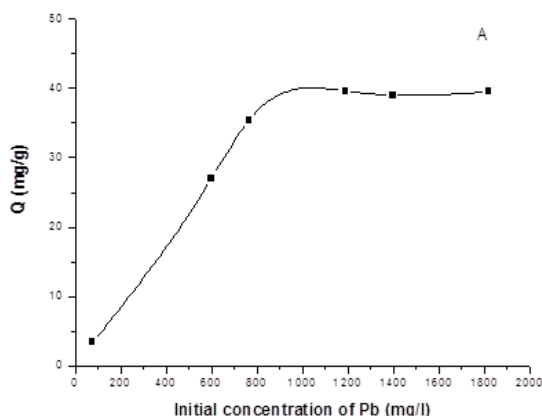


Fig. 1. Biosorption capacity of Pb(II) onto *nigella* biomass at varying initial concentration (A) and contact time (B)

Effect of contact time

The equilibrium time is one of the most important parameters for an economical wastewater treatment system. Fig. 1B shows the effect of contact time for the biosorption of Pb(II) by the *N. sativa* seeds biomass. The capacity of biosorption for Pb(II) was rapid in the initial stage (90–180 minutes). This is probably due to the biosorption of metal ions onto the *N. sativa* seeds exterior surface. In the subsequent stages, the biosorption became slow when the ions diffused into the interior surface. Furthermore, this can also be attributed to more vacant surface binding sites available for biosorption during the initial stage³². A maximum percentage metal removal of 80% and the capacity of biosorption of 90.31 mg/g was observed at 360 min. for Pb(II). Conversely, after a certain time, the capacity of biosorption for ion metal decreased when approaching 720 min. when biosorption sites became saturated and repulsive forces between the solute molecules on the solids came into play^{33, 34}. Similar results have also been reported previously for Pb(II) ions absorption into pomegranate peel³⁵ and tea waste³⁶. The contact time of 360 min. was used in subsequent tests of this study.

Effect of initial pH

The pH of medium in biosorption phenomenon is one of the pivotal parameters that affect the ionization of the surface functional groups and specification of metal ions and thus affecting the interactions of the metal ions at interface. In order to evaluate the adsorption of Pb(II) ions onto the *N. sativa* seeds biomass, the experiments were conducted at pH = 2, 4, 6, 8 and 10, while the biosorbent dose- contact time and initial metal

concentration were 0.07 g, 360 min. and 764 mg/L, respectively. The pH dependence of equilibrium adsorption data of Pb(II) ion is shown in Fig. 2 A. The pH profile shows that the removal of Pb(II) ion increase with solution pH and reaches the maximum value of 53.34% for Pb(II) at pH of 4. The sorption capacity (Q mg/g) increased until 23 mg/g. At low pH values, more protons compete with the metal cations for the biosorption sites as the biosorbent surface would be closely associated with hydronium ions³⁷. As pH increases, concentration of the hydronium ions decrease and the functional groups on the biosorbent surface become negatively charged, making the biosorbent surface more readily available for metal cations biosorption^{38,39}, thus increasing the biosorption capacity of the biosorbent. The fluctuations in biosorption yield were not only related to the formation of soluble hydroxylated complexes of the metal ions (lead ions in the form of $Pb(OH)_2$) but also to the ionized nature of the cell wall surface of the *N. sativa* seeds biomass under the studied pH. It can be speculated that isoelectric point of the biosorbent is approximately around a pH of 4. It was concluded that the biosorbent was active in a large spectra of pH, ranged from 4.0 to 7.0. In this study the pH of 4.5 was applied.

Effect of the biosorbent dose

In adsorption between sorbat and sorbent in a given medium, the dosage of biomass is an important parameter, which determines the capacity of uptake in function of initial concentration. The biosorption of Pb(II) onto *N. sativa* seeds as a function of the biosorbent dose was studied by varying the amount of biosorbent from 0.01 to 0.2 g/10 ml solution at pH 4.5 and 764 mg/L Pb(II) initial concentration for 360 min. contact time.

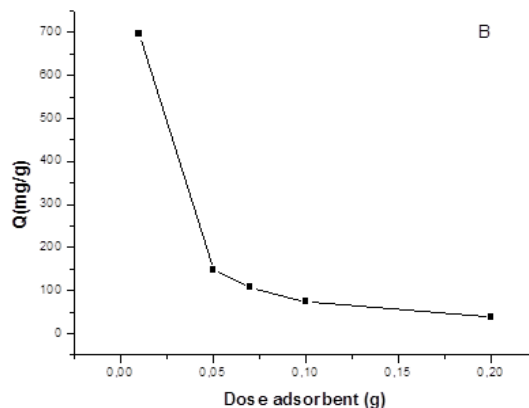
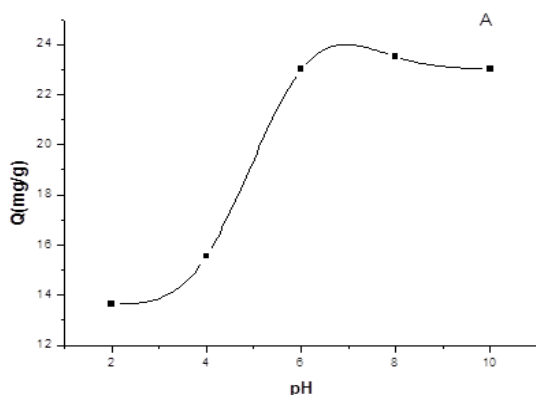


Fig. 2. Biosorption capacity of Pb(II) onto nigella biomass at various initial pH (A) and sorbent dosage (B)

The effect of the sorbent dose on the Q value (mg/g) is shown in Fig. 2B. At 0.01 g biosorbent dose for 10 ml of solution, the biosorption capacity of the *N. sativa* seeds biomass is high and tends to decrease at higher doses of biomass. This result demonstrates clearly that a very small dose of biomass was required to remove the provided amount of lead. In fact, this behavior may be due to the availability and accessibility of more adsorption sites at low dosage. At high biomass dosage the sites became less available due to the steric overlapping. In literature, similar results in agreement of ours were obtained on different biomass for the adsorption of Pb(II) in function of dosage^{40, 41}. Subsequently, the dose of 0.07 g was applied in the tests of this study.

Biosorption isotherms

To examine the relationship between adsorbed (Q_e) and equilibrium concentrations (C_e) of lead onto Nigella seeds biomass, sorption isotherm models of Langmuir and Freundlich⁴², which describe the equilibrium relationships between sorbent and sorbate in solution are applied. Sorption equilibrium is established when the concentration of sorbate in the bulk solution is in dynamic balance with that of the interface. To get the equilibrium data, the initial concentration of Pb(II) varied, while the biosorbent dose of biomass was kept constant. In Langmuir isotherm, the biosorption takes place at specific homogeneous sites of sorbent and was successfully applied in many process of monolayer biosorption. The linear form of the Langmuir isotherm is shown in equation 3.

$$\frac{C_e}{Q_e} = \frac{1}{KLQ_{max}} + \frac{C_e}{Q_{max}} \quad 3$$

Where Q_e is the metal ion sorbed (mg/g), C_e the equilibrium concentration of metal ion solution, and K_L is the Langmuir adsorption constant. A plot of C_e/Q_e versus C_e for the biosorption of Pb(II) onto the *N. sativa* seeds biomass gives a straight line (Fig. 3A). On the opposite, in Freundlich isotherm a heterogeneous empirical systems, which describe the phenomenon of adsorption at interface was employed. The log form of equation 4 was obtained by plotting $\log Q_e$ versus $\log C_e$ (Figure 3B).

$$\log Q_e = \frac{1}{n}(\log C_e) + \log K_f \quad 4$$

Where C_e (mg/L) is the equilibrium concentration, Q_e (mg/g) is the amount of metal ions adsorbed per specified amount of adsorbent at equilibrium; K_f and $1/n$ are the constants. The adsorption parameters evaluated from their isotherms (Fig. 3 A and B in shadow frame) demonstrate clearly that theoretical calculated adsorption capacity (Q_{max}) of Pb(II) at interface found in Langmuir model was in agreement with the experimentally determined sorption (39, 7mg/g). On the other hand, the correlation coefficient in Langmuir isotherm model is higher than the exhibited one by Freundlich. Thus, these two parameters suggest that the Langmuir isotherm describe better the sorption phenomenon than the Freundlich model.

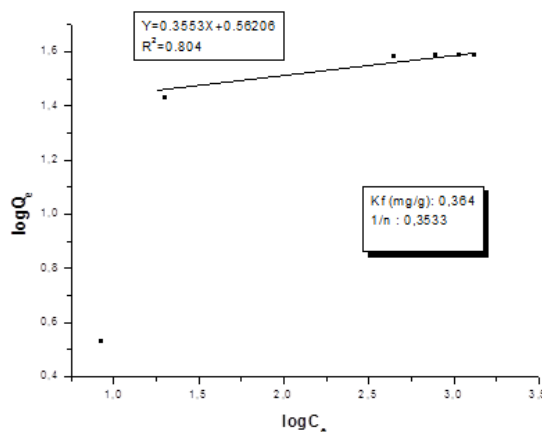
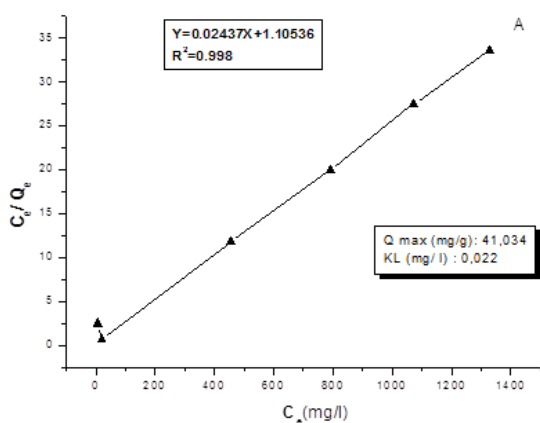


Fig. 3. Langmuir (A) and Freundlich (B) biosorption isotherms of Pb(II) onto nigella seeds biomass

Biosorption Kinetics

The rate controlling mass transfer, chemical reaction processes and biosorption mechanism can be effectively investigated by different kinetic models. In the present study, the pseudo-first-order and pseudo-second order models were used to determine the mechanism of the biosorption process. In the model of the pseudo-first-order given by the equation 5 was used in this study⁴².

$$\log(Q_e - Q_t) = \log Q_e - \frac{tK}{2303} \quad 5$$

Where Q_t is the amount of metal ions adsorbed (mg/g) at any arbitrary time t (min.), Q_e is the amount of metal ion adsorbed (mg/g) at equilibrium and K_1 is the pseudo-first order reaction rate constant for adsorption (min^{-1}) and t is the time (minutes). The plots of $\log(Q_e - Q_t)$ versus $\log Q_e$ of the pseudo-first-order model for Pb(II) biosorption on the *N. sativa* seeds biomass is shown in Fig. 4 A. For the pseudo-second-order model assumes that the rate of occupation of biosorption sites is proportional to the square of the number of unoccupied sites⁴². The linear form can be expressed from equation 6.

$$\frac{t}{Q_t} = \frac{1}{Q_e^2 K_2} + \frac{t}{Q_e} \quad 6$$

Where Q_e is the amount of metal ion adsorbed (mg/g) at equilibrium and K_2 is the rate constant of pseudo-second-order model ($\text{g mg}^{-1} \text{min}^{-1}$). The plots of (t/Q_t) versus t (min.) of the pseudo-second-order model for Pb(II)

biosorption on the *N. sativa* seeds biomass is shown in Figure 4B.

In view of the obtained of the kinetic parameters (in shadow frame), it appears that the uptake capacity in the pseudo-second order reflect well the nature of kinetic due to the similarity with the experimental value comparing to the pseudo-first order where the uptake represent only 56%. In effect, the correlation coefficient (R^2) value of the pseudo-second-order kinetic model is very close to one compared to the pseudo-first order that it could not used to describe the entire biosorption process. Consequently, it can be concluded that the biosorption process for Pb(II) ions at biomass interface followed a pseudo-second-order reaction.

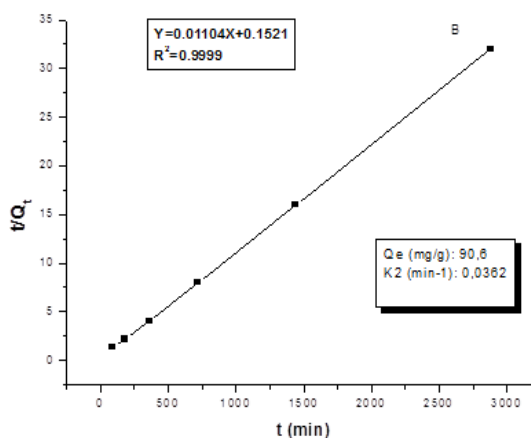
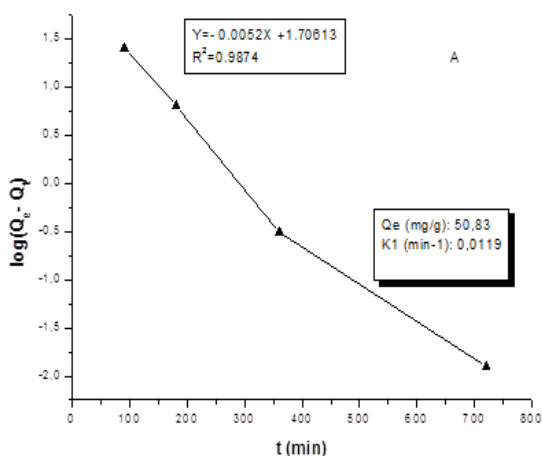


Fig. 4. Pseudo first order (A) and pseudo second one (B) kinetic models of sorption of Pb(II) onto nigella seeds biomass

Characterization of Biosorbent by FTIR

The FTIR spectra of before treatment and Pb(II) loaded *N. sativa* seeds are shown in Fig. 5. The native biomass of *N. sativa* seeds before sorption lead (Fig. 5 A) showed a broad band at 3404.36 cm^{-1} indicating the presence of $-\text{OH}$ group. The peak at 3008.95 cm^{-1} shows N-H stretching of amino acids as well as alkenes. The peak at 1745.58 cm^{-1} depicts carbonyl $\text{C}=\text{O}$ group. The peak at 1711.03 cm^{-1} shows aldehyde. The peak at 1654.92 cm^{-1} shows stretching vibration of C-OH band from proteins and alkenes.

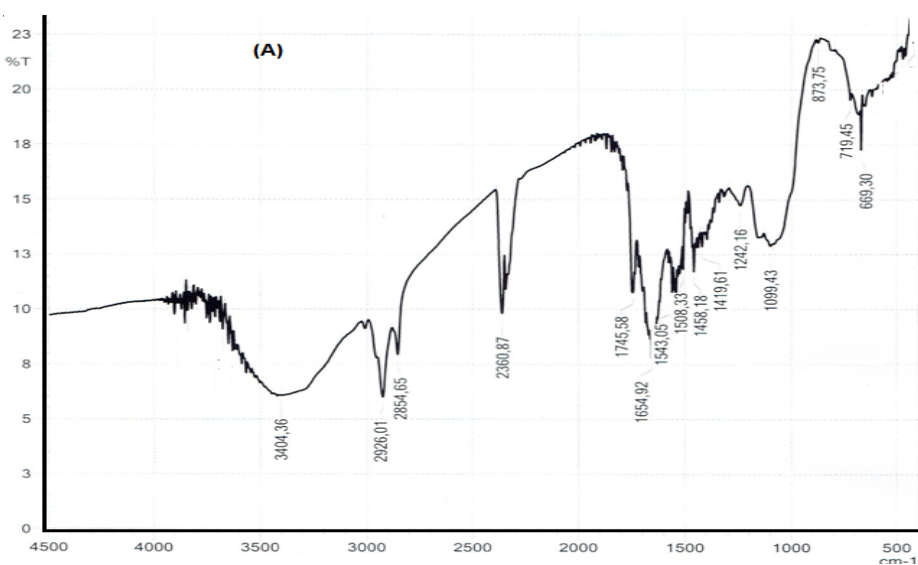


Fig. 5(A)

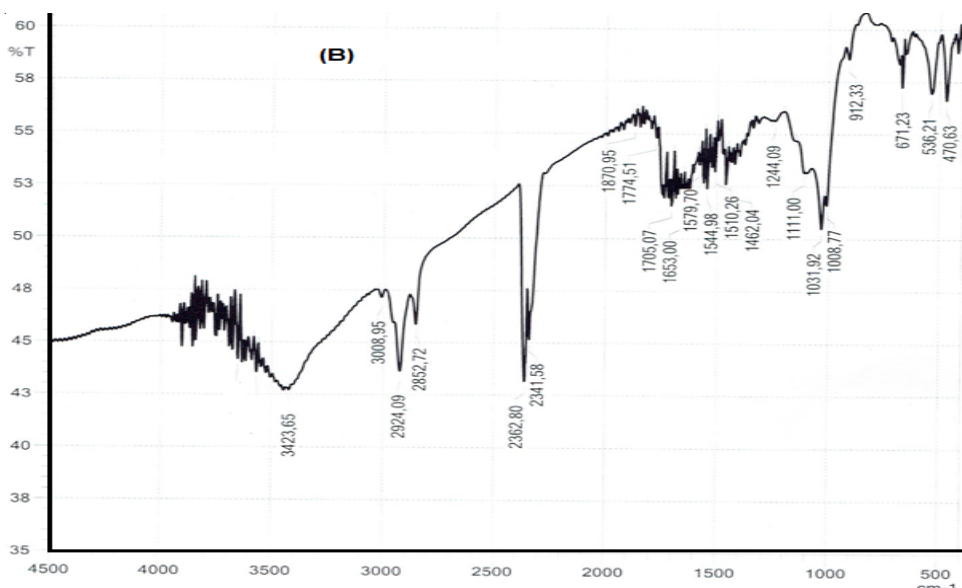


Fig. 5. FTIR spectra of *N. sativa* seeds biomass before (A) and after Pb (II) adsorption (B)

Efficiency removal of lead on real effluents

The obtained results of removing lead on the two studied effluents demonstrated a high efficiency sequestration of Pb(II). In fact, 97% of lead was removed in the sludge sample and 64% in decanted/filtrated effluent. Such results confirmed once more the efficient removal of Pb (II) heavy metal ion by *N. sativa* seeds biosorbent in aqueous systems. It can be concluded that *N. sativa* seeds biomass have a great potential for sustainable treatment of water and wastewater with respect to heavy metals.

CONCLUSION

Nigella sativa seeds biomass possesses an important sorption capacity towards Pb (II) under optimum conditions particularly pH medium, time contact and sorbent dose, which play a pivotal role in this phenomenon. The FTIR analysis indicates

the implication of hydroxyl and carbonyl groups in the interactions at interface. Langmuir isotherm model with an adsorption capacity of 41.034 mg/g and pseudo-second order kinetic with an r^2 close to 1 fitted well the experimental sorption data. The removal of Pb (II) from sludge that can attain 97% confirms the previous findings. So, this biomass is an effective sorbent for sorption of Pb(II).

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