



Phosphorus Removal and Recovery from High Phosphorus Wastewater by the HAP Crystallization Process

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ABSTRACT

In the experiment, the crystallization process was compared with the chemical precipitate process for phosphorus removal efficiency. Calcites were used as seeds to remove and recover phosphorus from synthetic phosphorite wastewater. The results indicated that under the same wastewater compositions, the crystallization process showed better phosphorus removal performance. At the initial stage, the corresponding removed Ca/P ratio (DCa/DP) was 1.43:1. When the phosphorus concentration was low, phosphorus could be removed further in the crystallization system, whereas the chemical precipitation stopped. During the reaction, calcium concentration firstly decreased because of precipitation or crystallization. Then it slightly increased, which possibly resulted from the transformation of HAP precursor or the dissolution of calcite seeds. The effects of pH, initial phosphorus concentration and Ca/P ratio were investigated. Phosphorus removal and recovery efficiency increased with the increment of the above three factors. In addition, the Response Surface Methodology was used to optimize the related factors. A significant quadratic model was obtained. From the optimization results, it could be found that when pH value was 9.0, molar ratio was 1.07 and reaction time was 36 h, the phosphorus removal rate could reach 96.1%.

Key words: calcite, hydroxyapatite, phosphorus recovery, phosphorus wastewater.

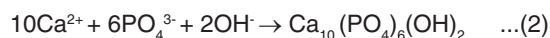
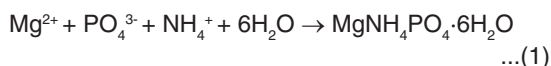
INTRODUCTION

Phosphorus is one of the limiting elements for eutrophication¹. An excessive discharged phosphorus in water bodies leads to algal blooms, resulting in the reduction of available oxygen and the deterioration of the water quality². The amounts of phosphorus in some types of wastewater are considerable, such as industrial wastewater³,

swine wastewater⁴, agricultural wastewater⁵, pharmaceutical wastewater⁶ and phosphorite wastewater⁷⁻⁸.

On the other hand, phosphorus is a nonrenewable resource and becoming rarer. And it has no substitute in the nature⁹. However the demand for phosphate rock is continually growing. According to the statistics from United States

Geological Survey, the phosphorus rock production in 2009 around the world was 166 million tons, which were 181 million tons in 2010 and 198 million tons in 2011¹⁰⁻¹². Consequently, much attention has been paid to phosphorus recovery from wastewater to realize the sustainable utilization of phosphorus. Struvite (MAP, MgNH_4PO_4) and hydroxyapatite (HAP, $\text{Ca}_5(\text{PO}_4)_3\text{OH}$) crystallization processes are the main methods. In the above process, crystals of MAP or HAP are formed by the following reactions.



By the above methods, some researchers have recycled phosphorus from various kinds of wastewaters, using different materials as seeds. Chen utilized xonotlite as the seed to recover phosphorus from the supernatant of concentrated sludge. The removal efficiency reached 91.3% after 24 h reaction¹³. Powdered converter slag was used as a seed material to recover phosphorus in the membrane-filtrate wastewater by Eung-Ho Kim¹⁴. High concentration of phosphorus was deposited on the surfaces of seed particles in the form of HAP. In this study, calcites were used as the seeds to remove and recover phosphorus from synthetic phosphorite wastewater.

As stated above, the phosphorus concentration in phosphorite wastewater is high. In order to recover phosphorus, in this experiment

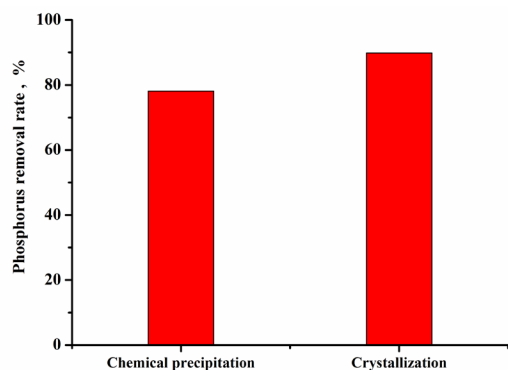


Fig.1: Phosphorus removal rates in crystallization and chemical precipitation systems after 72

chemical precipitate process was compared with crystallization process for the same synthetic phosphorite wastewater. Due to the low ammonia concentration, HAP was considered as the suitable recovery crystallization process. As the main factors, the effect of initial phosphorus concentration, pH value, molar ratio of calcium and phosphorus and reaction time on the removal efficiency were investigated. Furthermore, Design Expert software was used to optimize the related factors.

MATERIALS AND METHODS

Synthetic phosphorite wastewater

The phosphate concentration of the water samples was from 30 to 120 mg/L, and COD was about 100 mg/L. The wastewater was prepared by potassium dihydrogen phosphate (KH_2PO_4) and acetic acid (CH_3COOH). pH was adjusted with the solution of hydrochloric acid (HCl), sodium hydroxide (NaOH). Calcite was used as the seed.

Experimental procedure

The experiments were carried out in a constant-temperature air-bathing shaker at 25 °C with the rotational speed 120 r/min. The other parameters were listed in Table 1. The water samples were filtrated by 0.45µm membranes before analysis at different time intervals.

Analytical methods

pH was instrumentally measured by Sartorius PB-10 pH meter. Phosphorus was

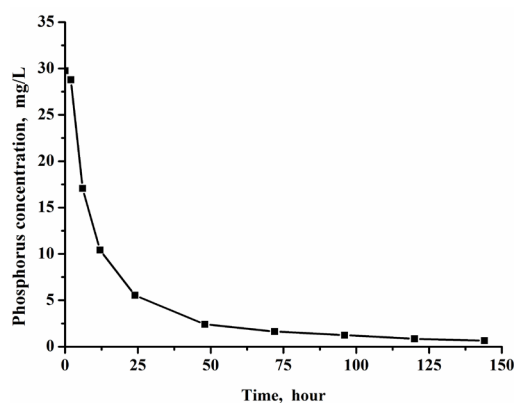


Fig. 2: The variations of phosphorus concentration (the initial phosphorus concentration was 60 mg/L, Ca/P ratio was 0.9, Seed dosage was 40 g/L, pH value was 9)

determined in accordance with ammonium molybdate spectrophotometric method¹⁵. And calcium ions were determined by an atomic absorption spectrometry (AAS) spectrometer. TOC was measured via a TOC analyzer (TOC-L-CPH, Shimadzu, Japan). The analyses of Response Surface Methodology (RSM) were used for parameter optimization by Design expert software.

RESULTS AND DISCUSSION

Comparison of induced crystallization and chemical precipitation

As shown in Fig.1, compared to chemical precipitation, crystallization obtained better phosphorus removal efficiency. At the initial stage, the two systems showed similar performance. After 6 hours, when the phosphorus concentration in the solution was low, in the crystallization system phosphorus could be removed continuously, whereas the removal was stopped in the precipitation system. The interface energy could be reduced by the calcite seeds, so phosphate and calcium were retained in the form of crystallization on seed surface. Moreover, in comparison with precipitation system, the product achieved by crystallization had less water content, which was easier to be separated and reused as fertilizer¹⁶⁻¹⁷. It is additional advantage of crystallization.

Variations of calcium and phosphorus in the induced crystallization system

In the induced crystallization system, the content of soluble phosphate decreased rapidly in 12h (Fig.2), declined with the reaction time and tended to a balance. Phosphorus removal efficiency reached almost 100%. The initial phosphorus concentration was 60 mg/L, but after calcium was

added the concentration it immediately dropped to 30 mg/L. Inevitably, calcium phosphorus precipitation was formed because of the supersaturation. About 30 mg/L phosphorus and 58 mg/L calcium were removed from the solution immediately after mixing, and the corresponding removed Ca/P ratio (DCa/DP) was 1.43:1. Consequently, it could be inferred that the possible formation of precipitate was $\text{Ca}_3(\text{PO}_4)_2$, with the Ca/P ratio of 1.5. When the phosphorus concentration was low, phosphorus could be removed further, which indicated that the induced crystallization could still occur whereas the chemical precipitation was stopped. At the low phosphorus and calcium concentrations, the interfacial energy was reduced by the calcite seed, so the phosphate crystals deposited on the surfaces of calcite. HAP could grow on under metastable state by the seed in the crystallization¹³.

Within 24 h, the average DCa/DP ratio was 1:5, much less than the DCa/DP ratio at the initial stage. In addition, the calcium had a slight increase after 24h. There were two possible reasons for this phenomenon. Firstly, the main form of calcite seed was calcium carbonate, which could provide calcium ions. Secondly, the precursor of HAP was transformed, making the Ca/P ratio change, and consequently partial of the calcium dissolved. The increased calcium ions were beneficial to restrain the phosphorus concentration in the solution.

The effect of pH on the recovery of phosphorus crystallization

It could be observed from Fig.4 that the efficiency of phosphate removal was improved at higher pH value. Furthermore, as shown in Fig.5, the removal processes was in accordance with pseudo-second-order kinetics equation during 0-48 h. The

Table 1: The parameters in the experiments

Factor	Phosphorus concentration / mg•L ⁻¹	Ca/P*	pH value
pH value	60	0.9	5,7,9
Ca/P*	60	0.6, 0.9, 1.2, 1.5, 1.8	9
Phosphorus concentration	30,60,90,120	0.9	9

*: mole ratio.

reaction rate constant increased with the increase of pH values. However, this did not mean that the higher pH could improve the crystallization quality. When pH was higher than 9, the bicarbonate ion would transform into carbonate ion, resulting in the calcium carbonate precipitate, mixed with the calcium phosphorus product¹³.

From Fig.6, it could be seen that when the pH was higher than 7, the calcium concentration reduced during 0~24 h, indicating the precipitation or crystallization occurred. But, at pH 5 the calcium increased due to the dissolution of the calcite seeds at weak-acid condition. After 24 h, the concentration of calcium increased slightly with the time. The concentration of TOC in the process of the whole reaction basically remained unchanged, indicating

that no organic matter was adsorbed on the seed surface, so the presence of organic matter would not affect the crystallization.

Effect of Ca/P ratio on the recovery of phosphorus crystallization

It can be seen from the Fig.7 that the Ca/P ratio had an impact on the recovery of crystallization. When the Ca/P ratio was 0.6, the removal rate of phosphorus was less than 90% within 96 h, which was significantly lower than the results of other Ca/P ratios. Although the better removal rate was achieved at higher Ca/P ratio, it was uneconomical to add excessive calcium for recovering phosphorus. High calcium dosage also resulted in the supersaturation, therefore the dominated reaction was calcium phosphorus precipitation, even though

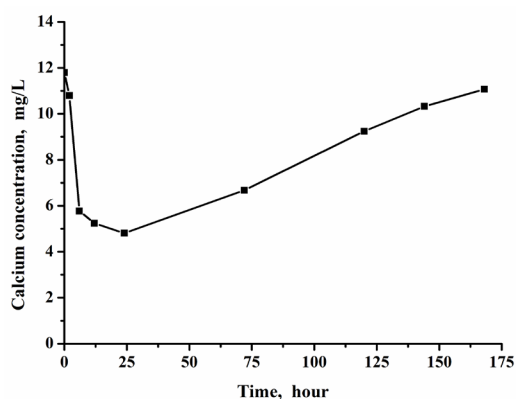


Fig. 3: The variations of calcium concentration (the initial phosphorus concentration was 60 mg/L, Ca/P ratio was 0.9, seed dosage was 40 g/L, pH value was 9)

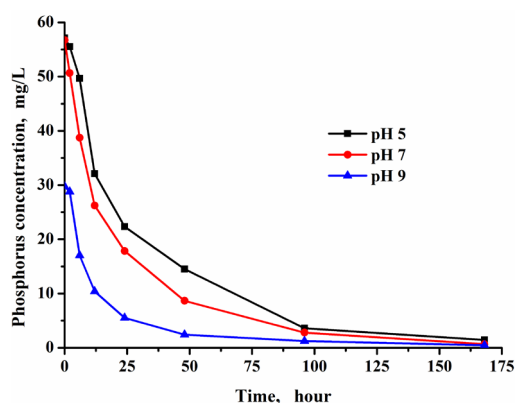


Fig. 4: The variations of phosphorus concentration at different pH values

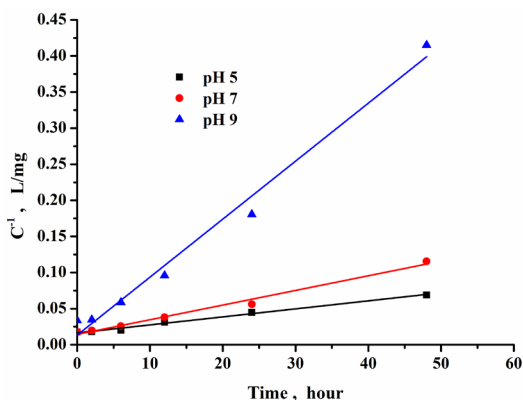


Fig. 5: The fitting of pseudo-second-order kinetics equation

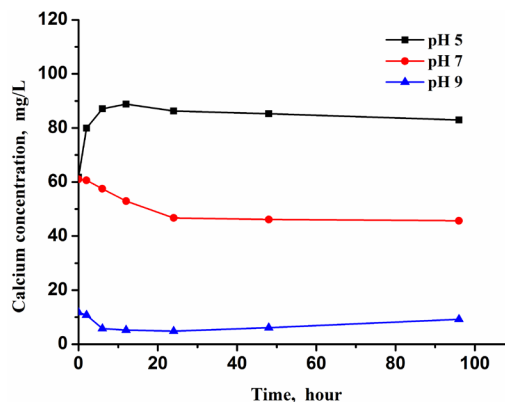


Fig. 6: The variations of calcium concentration at different pH values

the phosphorus removal efficiency increased. When the Ca/P ratio was 1.8, phosphorus was removed about 90% immediately after mixing, demonstrating that chemical precipitation was the main reason for the phosphorus removal. From an economical point of view, it is better to choose 0.9 as the Ca/P ratio. When the Ca/P ratio was 0.9, the recovery rate of phosphate can reached 96% within 48 h.

Effect of initial phosphorus concentration

It was indicated from the Fig.8 that under different initial concentrations of phosphorus, the phosphorus decreased rapid in the initial stage and leveled off over 48 h. pH values also changed during the reaction. Judging from the reaction mechanism, at the initial period, phosphorus was mainly removed with homogeneous precipitate.

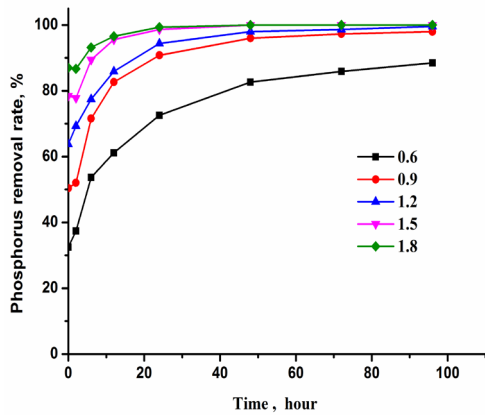


Fig. 7: The variations of phosphorus removal rate at different Ca/P ratios

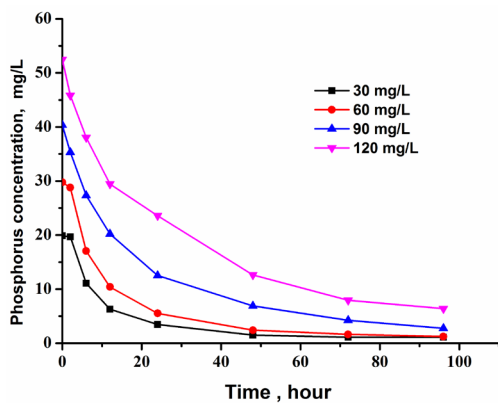


Fig.8: The variations of phosphorus removal rate at different initial phosphorus concentrations

The higher concentration of phosphorus in the solution, the more homogeneous precipitate occurred and the higher phosphorus removal rate was obtained. Then, the pH value increased, which possibly resulted from the transformation of HAP precursor and calcium carbonate coprecipitation. With the decrease of phosphorus concentration, the heterogeneous crystallization dominated gradually. As mentioned before, after 12 h, the concentration of calcium ion slightly increased, probably due to the dissolution from calcite seed. From this point, if more phosphorus wastewater were added, the phosphorus removal could continue.

Optimization by the response surface method

The process was optimized for the important parameters like pH, time, and Ca/P ratio. The removal rate of phosphorus was considered as the foundation to ascertain the best optimized factor which was directed by using Box-Behnken design. Box-Behnken design which belongs to the response surface methodology (RSM) was used in order to demonstrate the nature of the response and to clarify the optimum states of the most important independent variables¹⁸. Three major variables were time, pH and Ca/P ratio. These factors were tested at three different levels (low, basal, high) as shown in Table 2. According to the Box-Behnken design for three variables, a significant quadratic model (P-value < 0.0001) was obtained by using analysis of variance (ANOVA), which was listed as follows. The phosphorus removal rate = -283.89255 + 40.50968 (pH) + 221.17606 (Ca/P ratio) + 3.69905 (time) - 2.94167 (pH) × (Ca/P ratio) - 0.21958 (pH) × (time)

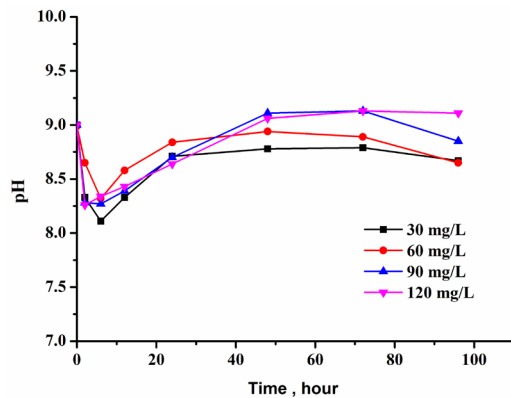
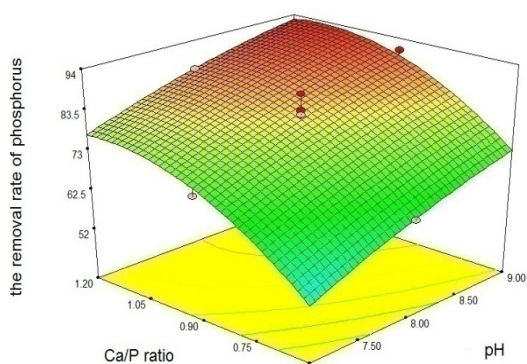
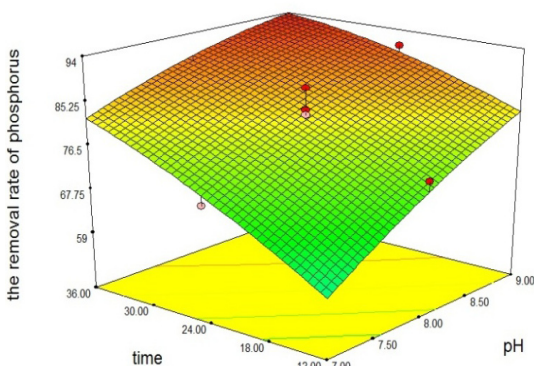


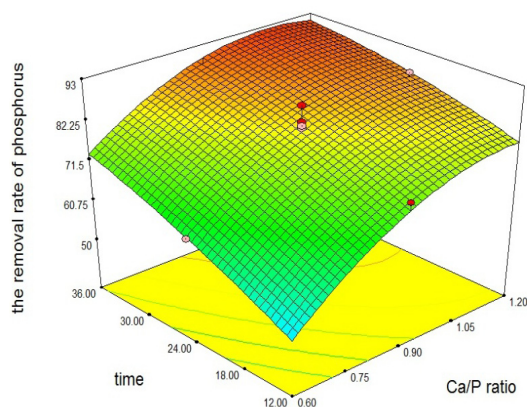
Fig. 9: The variations of pH at different initial phosphorus concentrations



(a) Surface plot of the removal rate of phosphorus (%) vs Ca/P ratio, pH



(b) Surface plot of the removal rate of phosphorus (%) vs time, pH



(c) Surface plot of the removal rate of phosphorus (%) vs time, Ca/P ratio

Fig. 10: 3D surface plots of the removal rate of phosphorus as a function: (a)Ca/P ratio and pH, (b) time and pH, (c) time and Ca/P ratio.

Table 2: The actual and predicted values of phosphorus removal rates

Standard order	pH	Ca/P	Time	Actual Value	Predicted Value
1	7	0.6	12	36.5	37.2
2	9	0.6	12	61.1	61.4
3	7	1.2	12	67.2	66.7
4	9	1.2	12	85.9	87.5
5	7	0.6	36	67.0	65.2
6	9	0.6	36	78.6	79.0
7	7	1.2	36	84.9	84.5
8	9	1.2	36	95.5	94.7
9	7	0.9	24	70.3	72.2
10	9	0.9	24	90.8	89.4
11	8	0.6	24	63.4	63.9
12	8	1.2	24	86.4	86.5
13	8	0.9	12	73.9	71.8
14	8	0.9	36	86.8	89.5
15	8	0.9	24	80.8	82.3
16	8	0.9	24	81.0	82.3
17	8	0.9	24	87.1	82.3
18	8	0.9	24	81.8	82.3
19	8	0.9	24	82.8	82.3
20	8	0.9	24	81.8	82.3

$$-0.71806 (\text{Ca/P ratio}) \times (\text{time}) - 1.49864 (\text{pH})^2 - 79.26263 (\text{Ca/P ratio})^2 - 0.011692 (\text{time})^2.$$

The corresponding analysis of variance (ANOVA) showed the variables that were extremely significant were the square term of Ca/P ratio. Furthermore, the linear effect of pH, and Ca/P ratio and time were significant at the level of p-value 0.001. The 3D plots for the relation between two variables were illustrated in Fig.10. From the optimization results, when pH value was 9.00, Ca/P ratio was 1.07 and reaction time was 36 h, the phosphorus removal rate can reach 96.1%.

CONCLUSION

In the experiment, the crystallization process was compared with the chemical precipitate process for phosphorus removal efficiency. Under the same wastewater compositions, the crystallization process showed better phosphorus removal performance. During the crystallization reaction,

calcium concentration firstly decreased because of precipitation or crystallization. Then it slightly increased, which possibly resulted from the transformation of HAP precursor or the dissolution of calcite seeds. Phosphorus removal and recovery efficiency increased with the increment of the pH, initial phosphorus concentration and Ca/P ratio. Moreover, a significant quadratic model was obtained by the Response Surface Methodology. The optimization results showed that when pH value was 9.0, molar ratio was 1.07 and reaction time was 36 h, the phosphorus removal rate could reach 96.1%.

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REFERENCES

1. Oladoja, N. A.; Olojede, I. A.; Adesina, A.O.; Adelagun R. O. A.; Sani Y. M. *Chem. Eng. Res. Des.* **2013**, *91*, 810-818
2. Lüring, M.; Oosterhout, F.V. *Water. Res.* **2013**, *47*, 6527-6537
3. Huang, Y H; Shih, Y. J.; Chang, C. C.; Chuang, S. H. *Desalin. Water Treat.* **2011**, *32*, 351-356
4. Suzuki, K.; Tanaka Y.; Kuroda, K.; Hanajima, D.; Fukumoto, Y. *Bioresour. Technol.* **2005**, *96*, 1544-1550
5. Yi, W. G.; Lo, K. V. *J. Environ. Sci. Health., Part B.* **2003**, *38*, 501-509
6. Qiu, G. L.; Song, Y. H.; Zeng, P.; Xiao, S.; Duan, L. *Chemosphere.* **2011**, *84*, 241-246
7. Xu, C. W.; Gao, H. L. *Environ. Technol.* **2004**, *4*, 25-28
8. Lu, Y. Research on Treatment and Reuse Technic of Phosphorite Processing wastewater. PhD thesis, Wuhan University of Technology, Wuhan, **2010**.
9. Shu, L.; Schneider, P.; Jegatheesan, V.; Johnson, J. *Bioresour Technol.* **2006**, *97*, 2211-2216
10. U.S. Geological Survey. Mineral Commodity Summaries 2011. U.S. Government Printing Office, Washington, **2011**.
11. U.S. Geological Survey. Mineral Commodity Summaries 2012. U.S. Government Printing Office, Washington, **2012**.
12. U.S. Geological Survey. Mineral Commodity Summaries 2013. Washington U.S. Government Printing Office, Washington, **2013**.
13. Chen, X. C.; Kong, H. N.; Wu, D. Y.; Wang, X. Z.; Lin, Y.Y. *J. Environ. Sci.* **2009**, *21*, 575-580
14. Kim, E. H.; Yim, S. B.; Jung, H. C.; Lee, E. J. *J. Hazard. Mater.* **2006**, *136*, 690-697
15. American Public Health Association, American Water Works Association and Water Environment Federation. Standard Methods for the Examination of Water and Wastewater. American Standard, **2006**.
16. Mañas, A.; Biscans, B.; Spérandio, M. *Water. Res.* **2011**, *45*, 3776-3786
17. De-Bashan, L. E.; Bashan, Y. *Water. Res.* **2004**, *38*, 4222-4246
18. Box, G. E.; Wilson, K. B. *J. R. Statist. Soc.* **1951**, *13*, 1-45